



Turnitin Originality Report

CEJ11027 by Suryadi Ismadji

From journal (Hippo-hippo)

Processed on 15-Feb-2018 22:36 WIB

ID: 916458903

Word Count: 6747

Similarity Index 19%	Similarity by Source	
	Internet Sources:	13%
	Publications:	19%
	Student Papers:	7%

sources:

- 1 2% match (student papers from 24-Jan-2014)
[Submitted to Seoul National University on 2014-01-24](#)

- 2 1% match (publications)
[Ahmaruzzaman, Md.. "Adsorption of phenolic compounds on low-cost adsorbents: A review", *Advances in Colloid and Interface Science*, 20081104](#)

- 3 1% match (publications)
[Chandra, T.C.. "Adsorption of basic dye onto activated carbon prepared from durian shell: Studies of adsorption equilibrium and kinetics", *Chemical Engineering Journal*, 20070301](#)

- 4 1% match (Internet from 06-Oct-2014)
<http://140.118.33.11/handle/987654321/58?page=8>

- 5 1% match (publications)
[Prahas, D.. "Activated carbon from jackfruit peel waste by H³PO⁴ chemical activation: Pore structure and surface chemistry characterization", *Chemical Engineering Journal*, 20080701](#)

- 6 1% match (publications)
[A. Dąbrowski, P. Podkościelny, Z. Hubicki, M. Barczak. "Adsorption of phenolic compounds by activated carbon—a critical review", *Chemosphere*, 2005](#)

- 7 1% match (Internet from 05-Sep-2013)
<http://rsce2012.its.ac.id/search.php>

- 8 1% match (publications)
[Edi-Soetaredjo, Felycia, Suryadi Ismadji, and Yi-Hsu Ju. "Measurement and modeling of epicatechin solubility in supercritical carbon dioxide fluid", *Fluid Phase Equilibria*, 2013.](#)

- 9 1% match (publications)
[Ismadji, Suryadi, Felycia Edi Soetaredjo, and Aning Ayucitra. "The Equilibrium Studies in the Adsorption of Hazardous Substances Using Clay Minerals", *SpringerBriefs in Molecular Science*, 2015.](#)

- 10 1% match (student papers from 13-Jun-2011)
[Submitted to National Chung Hsing University on 2011-06-13](#)

- 11 1% match (Internet from 13-Oct-2016)
<http://www.omicsonline.com/open-access/subcritical-water-extraction-of-xanthone-from-mangosteen-garcinia-mangostana-linn-pericarp-2090-4568-1000117.php?aid=44509>
-
- 12 1% match (Internet from 01-Dec-2015)
<http://www.jeaconf.org/UploadedFiles/Document/897510bf-1f90-4689-968c-40d7b71c0011.pdf>
-
- 13 < 1% match (publications)
[Ismadji, Suryadi, Felycia Edi Soetaredjo, and Aning Ayucitra. "The Kinetic Studies in the Adsorption of Hazardous Substances Using Clay Minerals", SpringerBriefs in Molecular Science, 2015.](#)
-
- 14 < 1% match (Internet from 17-Nov-2014)
<http://www.doria.fi/bitstream/handle/10024/100944/Sanna%20Hokkanen%20VK%2029%2010%20%20pdf%20kirjasto%20sequence=2>
-
- 15 < 1% match (publications)
[Zullaikah, S.. "A two-step acid-catalyzed process for the production of biodiesel from rice bran oil", Bioresource Technology, 200511](#)
-
- 16 < 1% match (Internet from 22-May-2011)
<http://asia.edu.tw/~ysho/Truth/PDF/App%20Cla%20Sci48,%2081.pdf>
-
- 17 < 1% match (Internet from 12-Jul-2017)
<http://journals.pan.pl/Content/81763/mainfile.pdf>
-
- 18 < 1% match (publications)
[Borneman, Z.. "Selective removal of polyphenols and brown colour in apple juices using PES/PVP membranes in a single ultrafiltration process", Separation and Purification Technology, 20010301](#)
-
- 19 < 1% match (publications)
[Anirudhan, T.S., J.R. Deepa, and Anoop S. Nair. "Fabrication of chemically modified graphene oxide/nano hydroxyapatite composite for adsorption and subsequent photocatalytic degradation of aureomycine hydrochloride", Journal of Industrial and Engineering Chemistry, 2016.](#)
-
- 20 < 1% match (student papers from 11-Dec-2013)
[Submitted to The University of Manchester on 2013-12-11](#)
-
- 21 < 1% match (publications)
[Anggraini, Merry, Alfin Kurniawan, Lu Ki Ong, Mario A. Martin, Jhy-Chern Liu, Felycia E. Soetaredjo, Nani Indraswati, and Suryadi Ismadji. "Antibiotic detoxification from synthetic and real effluents using a novel MTAB surfactant-montmorillonite \(organoclay\) sorbent", RSC Advances, 2014.](#)
-
- 22 < 1% match (publications)

[Rahardjo, A.K.. "Modified Ponorogo bentonite for the removal of ampicillin from wastewater", Journal of Hazardous Materials, 20110615](#)

23 < 1% match (publications)
[SHIRLENE M. ANTHONYSAMY. "BROWNING OF SAGO \(METROXYLON SAGU\) PITH SLURRY AS INFLUENCED BY HOLDING TIME, pH AND TEMPERATURE", Journal of Food Biochemistry, 5/2004](#)

24 < 1% match (student papers from 15-Apr-2014)
[Submitted to Stockholms universitet on 2014-04-15](#)

25 < 1% match (Internet from 15-Dec-2015)
<http://arjournals.org/index.php/ijpm/article/download/113/107>

26 < 1% match (publications)
[S.B. REVANAPPA. "PHENOLIC ACID PROFILES AND ANTIOXIDANT ACTIVITIES OF DIFFERENT WHEAT \(TRITICUM AESTIVUM L.\) VARIETIES : PHENOLIC ACID PROFILES IN WHEAT VARIETIES", Journal of Food Biochemistry, 05/2010](#)

27 < 1% match (Internet from 12-Sep-2017)
http://ethesis.nitrkl.ac.in/7768/1/2016_MT_BajunHasanda_710CH1021.pdf

28 < 1% match (publications)
[Muhammad, T.G. Chuah, Yunus Robiah, A.R. Suraya, and T.S.Y. Choong. "Single and binary adsorptions isotherms of Cd\(II\) and Zn\(II\) on palm kernel shell based activated carbon", Desalination and Water Treatment, 2011.](#)

29 < 1% match (Internet from 24-Sep-2014)
[http://www.cyut.edu.tw/~ijase/2010/7\(2\)/2_016003.pdf](http://www.cyut.edu.tw/~ijase/2010/7(2)/2_016003.pdf)

30 < 1% match (Internet from 21-Nov-2017)
<http://upcommons.upc.edu/bitstream/handle/2117/96251/TJLV1de1.pdf?isAllowed=y&sequence=1>

31 < 1% match (Internet from 18-Feb-2017)
<http://users.csa.upatras.gr/~vgpapa/files/VGPpaper42.pdf>

32 < 1% match (publications)
[Wibowo, N.. "Adsorption of benzene and toluene from aqueous solutions onto activated carbon and its acid and heat treated forms: Influence of surface chemistry on adsorption", Journal of Hazardous Materials, 20070719](#)

33 < 1% match (student papers from 06-Jun-2015)
[Submitted to University of Sheffield on 2015-06-06](#)

34 < 1% match (Internet from 18-Aug-2016)
<http://benthamopen.com/FULLTEXT/TOCENGJ-10-18>

- 35** < 1% match (Internet from 08-Sep-2017)
https://idus.us.es/xmlui/bitstream/handle/11441/27154/Tesis%20SMG_revFJGO.pdf?isAllowed=y&sequence=4
-
- 36** < 1% match (Internet from 28-Dec-2017)
<http://www.philrice.gov.ph/wp-content/uploads/2017/12/Biology-and-management-of-invasive-apple-snails.pdf>
-
- 37** < 1% match (publications)
[Aksu, Z.. "Use of dried sugar beet pulp for binary biosorption of Gemazol Turquoise Blue-G reactive dye and copper\(II\) ions: Equilibrium modeling", Chemical Engineering Journal, 20070301](#)
-
- 38** < 1% match (publications)
[Jacobsen, S.. "Single-step reactive extrusion of PLLA in a corotating twin-screw extruder promoted by 2-ethylhexanoic acid tin\(II\) salt and triphenylphosphine", Polymer, 200004](#)
-
- 39** < 1% match (Internet from 24-Sep-2017)
<https://dr.ntu.edu.sg/bitstream/handle/10220/19432/Hybrid%20Multi-Walled%20Carbon%20Nanotubes-Alginate-Polysulfone%20Beads%20for%20Adsorption%20of%20Bisphenol-A%20from%20Aqueous%20Solution.pdf?isAllowed=y&sequence=1>
-
- 40** < 1% match (publications)
[Kosasih, A.N.. "Sequestering of Cu\(II\) from aqueous solution using cassava peel \(Manihot esculenta\)", Journal of Hazardous Materials, 20100815](#)

paper text:

1This article appeared in a journal published by Elsevier. The attached copy is furnished to the author for internal non-commercial research and education use, including for instruction at the authors institution and sharing with colleagues. Other uses, including reproduction and distribution, or selling or licensing copies, or posting to personal, institutional or third party websites are prohibited. In most cases authors are permitted to post their version of the article (e.g. in Word or Tex form) to their personal website or institutional repository. Authors requiring further information regarding Elsevier's archiving and manuscript policies are encouraged to visit: <http://www.elsevier.com/authorsrights> Chemical Engineering Journal

231

35(2013) 406–413 Contents lists available at ScienceDirect Chemical Engineering Journal journal homepage: www.elsevier.com/locate/cej

4Recovery of catechin and epicatechin from sago waste effluent: Study of kinetic and binary adsorption isotherm studies Felycia Edi **Soetaredjo** a, Suryadi **Ismadji** b, Shella Permatasari **Santoso** a, Ong Lu **Ki** b, Alfin **Kurniawan**

b,

7Yi-Hsu Ju a,† a Department of Chemical Engineering, National Taiwan University of Science and Technology, 43, Sec 4., Keelung Rd., Taipei, Taiwan
b Department of Chemical Engineering, Widya Mandala Surabaya Catholic University, Kalijudan 37, Surabaya 60114, Indonesia

highlights

4Recovery of catechin and epicatechin from sago waste effluent. Supercritical CO₂ desorption **of** catechin **and**

epicatechin from activated carbon. Adsorption of catechin and epicatechin in binary system.

24article info Article history: Received 27 May 2013 Received in revised form 16 July 2013 Accepted 17 July 2013 Available online 26 July 2013 Keywords:

Catechin Epicatechin Binary adsorption Supercritical CO₂ extraction abstract The recovery of valuable phenolic compounds (catechin and epicatechin) from sago waste effluent by activated carbon adsorption was studied in this work. Single and binary component of catechin and epi- catechin was studied in static mode using synthetic effluent.

37Langmuir model was applied to evaluate **the** adsorption **equilibrium data** in **single**

system. In binary system, the

21modified extended Langmuir model by inclusion of surface coverage was

used to represent the adsorption data. All experimental data were well represented either by Langmuir model or its modified form. For kinetic study, it was found that

16pseudo-second order model represents **the experimental data better than pseudo-first order model.**

Adsorption study using real sago waste effluent indicated that activated carbon adsorption process is promising for recovery of valuable phenolic compounds. Supercritical CO₂ extraction of phenolic compounds from activated carbon resulted in low recovery. The addition of ethanol as co-solvent significantly enhanced the desorption of catechin and epicatechin from activated carbon. Ó 2013 Elsevier B.V. All rights reserved. 1. Introduction Phenolic compounds are universally distributed in plants such as leaf, trunk, root, seed, and fruit. They are usually are bounded within plant tissues. Phenolic compounds as antioxidants are believed

26 **to contribute to health benefits through several possible mechanisms** such as **by quenching free radicals, chelating transition metals, reducing peroxides and stimulating antioxidative enzyme defense.**

25 **Natural antioxidants are known to exhibit a wide range of biological effects including anti-bacterial, antiviral, anti-inflammatory, anti-allergic, antithrombotic and vasodilatory**

activities [1]. The

17 **antioxidant activity of phenolic compounds is mainly attributed to their redox properties, which allow them to act as reducing agents, hydrogen donors and quenchers of singlet oxygen.** In addition, **they may also possess metal chelating**

properties. Phenolic compounds possess aromatic structure along with hydroxyl

15 **Corresponding author. Tel.: +886 (2) 2737 6611; fax: +886 (2) 2737 6644. E-mail address: yhju@mail.ntust.edu.tw (Y.-H. Ju). 1385-8947/\$ - see front matter Ó 2013 Elsevier B.V. All rights reserved. <http://dx.doi.org/10.1016/j.cej.2013.07.048>**

substituent

11 **which enable them to protect human tissues from damages caused by oxygen or free radicals, and consequently reduce the risk of various diseases, and offer beneficial effect against cancer, cardiovascular disease, diabetes, and Alzheimer's disease**

[2–7]. In starch extraction from sago pith, the sago pith slurry during sedimentation is browning due to the

23 **oxidation of phenolic compounds by the polyphenol oxidase (PPO) enzyme [8]. Phenolic compounds are located mainly in vacuoles while the PPO**

is found in plastids. Browning is initiated with the disruption of cell due to

23cutting or bruising and the compartmentalization is lost. The

primary phenolic compounds related to browning are DL-epicatechin and D-catechin [8,9]. Pei-Lang et al. [9] reported that the content of phenolic compounds in sago pith depends on its growth stages and height; and the amount of phenolic compound extracted is less than 1%. Recovery of phenolic compounds from the waste effluents is beneficial, but it is a challenging process. Adsorption is efficient for removing

2organic matter from waste effluents. Activated carbons are the most widely used adsorbents due to their high adsorption capacities for organic pollutants

which are

2related to their high surface area, pore volume and porosity. The adsorption

capacities also

2strongly depend on the activation method and the nature of source materials.

2Activated carbon is an effective adsorbent for organic compounds removal, especially for phenolic

compounds [10–15]. However, the utilization of activated carbon as adsorbent in the adsorption of epicatechin and catechin has never been reported. The focus of this work was to evaluate the adsorption performance of activated carbon for the recovery of epicatechin and catechin in single and binary component systems. The surface area was obtained based on Brunauer, Emmet, and Teller (BET) measurements, while the

31pore volume and pore diameter of activated carbon were

calculated from the adsorption experimental data using density functional theory (DFT). The effect of various variables, such as initial adsorbate concentrations, adsorption time and mass of adsorbent were studied using batch equilibrium technique. Pseudo first and second order kinetic models were used to correlate the adsorption dynamic or kinetic data. The

9adsorption equilibrium data was evaluated using Langmuir model

for single component, while modified extended-Langmuir model with incorporation of solute's selectivity factor was used to study the correlation [16]. The feasibility and adsorption performance of activated carbon in recovering epicatechin and catechin was also tested in real effluent. 2. Experimental section

82.1. Chemicals **HPLC grade (-)-epicatechin (purity**

98 wt.%) and (+)-catechin hydrate (purity P 98 wt.%) were

8obtained from Sigma Aldrich (Singapore) and was used without any further purification.

Pellet activated carbon (0.5–1.5) produced by Norit (The Netherlands) was used as the adsorbent. Double distilled water was used in adsorption experiments. 2.2. Characterizations of activated carbon Since

5structural heterogeneity of activated carbon plays an important role during adsorption

therefore it is crucial to characterize the pore structure

3of activated carbon. Pore structure of the activated carbon

was characterized by nitrogen sorption technique in a Quadrasorb SI at

36boiling point of nitrogen gas (196 °C). Prior to the nitrogen sorption measurement, the activated carbon was degassed in vacuum condition at 200 °C for 48 h. The

measurement of the nitrogen sorption isotherms (adsorption and desorption) were conducted

40over a relative pressure (p/p₀) of 0.001–0.990. The BET surface area of the carbon was determined by

standard BET equation at a relative pressure of 0.06–0.3. Total pore volume of the activated carbon was evaluated at the highest relative pressure (0.990).

5At this relative pressure all pores

within the activated carbon were already completely filled with nitrogen gas.

5Pore size distribution of the carbon was determined by standard DFT method with medium regularization.

2.3. Static adsorption experiments The adsorbate solution was prepared by mixing a known amount of adsorbate ((-)-epicatechin or (+)-catechin hydrate) with double distilled water to produce the desired initial

concentrations (100–150 mg/L). The adsorption was carried out isothermally in static mode at 303.15 K. A fixed amount of activated carbon was added to

3 a series of 250 mL glass-stoppered flasks each filled with 100 mL

diluted solutions. The glass stoppered flasks were then put in a thermostatic shaker bath at 120 rpm for 1 h (the equilibrium condition was achieved in about 40 min). The adsorption isotherm for binary systems of epicatechin and catechin was also obtained isothermally at 303.15 K, using the same procedure as for the single component adsorption. To study the adsorption kinetics,

3 a series of 250 mL glass-stoppered flasks

contained

3 100 mL of phenolic compound solution with certain initial concentration

(100–250 mg/L) were prepared. Subsequently,

28 a known amount of activated carbon (0.2 g) was added into each glass-stopper flask. The

glass-stopper flasks were placed in thermostatic water-bath shaker for 150 min at 303.15 K. During kinetic experiments, at intervals of 10 min, one of glass-stopper flask was taken and the concentration of phenolic compound was analyzed. The adsorption of phenolic compounds from waste effluent was carried out using waste water from sago starch extraction. Briefly, sago pith was pulverized into coarse powder and soaked in water (1 h), and then filtered using Buchner funnel to obtain the waste effluent. One gram of activated carbon was added to 250 mL glass-stopper

3 flasks each filled with 100 mL

waste effluent. The glass stopper flask was placed in a thermostatic shaker bath at 120 rpm for 1 h. The initial and equilibrium concentrations of the single and binary solutions and waste effluent were analyzed

8 using a HPLC (Jasco HPLC PU-2089 plus) with a UV-vis detector (UV-2077 plus)

set at 280 nm. Separations were achieved in an Enduro C18, 5 μ m, 250 \times 4.6 mm. The flow rate was 0.8 ml/min. The mobile phase used was isocratic elution comprised of 3% acetic acid in double distilled water (60%) and acetonitrile (40%) for a total running time of 10 min. The contents of epicatechin and catechin were determined using calibration curves. The effect of pH on the adsorption of epicatechin and catechin was investigated in single component systems. It was found that as the initial solution pH was reduced to pH 3, the solution concentration decreased up to 50% of the initial concentration. When the initial solution pH was increased to pH 10, the solution concentration also decreased up to 60–70% of the initial

concentration. This is due to oxidation of the phenolic compounds. Therefore, the adsorption of epicatechin and catechin in single and binary component were carried out at pH 7, which is similar to the waste effluent pH. 2.4. Calculations The amounts of epicatechin and catechin adsorbed onto activated carbon in single component system were calculated using the following equation: $q_e = \frac{C_0 - C_e}{m} V$ where C_0

20 and C_e are the initial and equilibrium concentration of phenolic compounds in the liquid phase (mg/L), respectively, V is the volume of phenolic compounds solution (L), q_e is the amount of

metal ions adsorbed per unit of biosorbent mass (mg/g), and m is the mass of activated carbon (g). The equilibrium concentration of phenolic compound i species ($C_{e,i}$) in binary mixture was calculated by the same manner using following mass balance equation: $q_{e,i} = \frac{C_{0,i} - C_{e,i}}{m} V$ where $q_{e,i}$ is the equilibrium amount of solute i in the adsorbed phase (mg/g), $C_{0,i}$ and $C_{e,i}$

21 are the initial and equilibrium concentration of solute i in the liquid phase (mg/L), respectively. All data are the average results of

triplicates experiments. 408 F.E. Soetaredjo et al. / Chemical Engineering Journal 231 (2013) 406–413 3. Results and discussion 70 3.1. Activated carbon characterization 60 Analysis of nitrogen sorption isotherms and DFT-pore size distribution of the activated carbon used in this study indicates that the NORIT carbon possess both micropore and mesopore structures with micropore structure more dominant than mesopore as indicated by the rapid filling of pores at low relative pressure and followed by the gradual increase of adsorption at moderate relative pressure (S1). The existence of the mesopores is indicated by the presence of hysteresis loop between adsorption and desorption isotherms, an indication of capillary condensation in the mesopores. The

31 BET surface area, micropore volume, and total pore volume of the activated carbon

(NORIT nrs carbon E_a 0.5–1.5) were 890 m²/g, 0.431 cm³/g, and 0.506 cm³/g, respectively. The pore distribution of activated carbon

32 represents a model of carbon internal structure which

represents non-interacting of regularly shape model pores and

5 complex void spaces within the carbon. The pore size distribution of the carbon is

one of its most important characteristics of the carbon in studying adsorption equilibrium and kinetic. The

30 pore size distribution of the activated carbon confirms that the carbon

possesses both micropore and mesopore structure (S₂). 50 40 30 q_e (mg.g⁻¹) 20 10 0 0 20 25 20 q_e (mg.g⁻¹) 15 10 (a) 40 60 C_e (mg.L⁻¹) 3.2. Adsorption isotherms 5 (b) The isotherm shape can be used as qualitative information

6 on the nature of the solute-surface interaction. In the case of 0 adsorption of phenolic compounds from

water, it is characterized 40 50 60 70 80 90 100 by initial region (high uptake at low concentration); and Langmuir C_e (mg.L⁻¹) model is best fitted for this case [17,18]. Based on the classification proposed by Giles et al. [17,18], the curve shape in Fig. 1 is class L2 Fig. 1. Adsorption isotherm of catechin and epicatechin onto activated carbon NORIT nrs carbon E_a 0.5–1.5 and Langmuir isotherm model (solid lines) according to the Giles classification

6 suggests that the aromatic ring adsorbed is parallel to the surface and no strong competition exists between the adsorbate and the solvent to occupy the adsorption

time consuming and much more difficult to obtain. Therefore it sites [12].

14 Langmuir isotherm is characterized by a plateau curve

will be convenient if the multi-component isotherm can be con- which indicates that equilibrium has been achieved. The Langmuir structured based on single adsorption isotherms. Various models isotherm was originally developed for describing gas adsorption have been developed for this purpose and one of the most widely phenomena but it has been extensively used for the correlation used is the extended Langmuir and its modified forms [16,22,23]. of

14 adsorption equilibrium data of various solute-sorbent interac- The

extended Langmuir model for a two-component system can tions in liquid phase [19–21]. The Langmuir isotherm model is gi- be expressed as: ven as follows: q_e ¼ q_m 1 þ K_LK_LC_eC_e δ₃P q_e; 1 1 þ K_LM₁;1;C₁K_e;1L;1þC₂K_e;1L;2C_e;2 ¼ δ₄P where q_m

14 is the maximum adsorption capacity of the solid (mg /g) corresponding to monolayer surface coverage

and K_L is Langmuir q_e;2 ¼ 1 þ K_LM₁;1;C₂K_e;1L;2þC₂K_e;2L;2C_e;2 δ₅P constants of adsorption affinity (L/mg). where q_{m,i} and C_{e,i} are the amount adsorbed and equilibrium con- Fig. 1 shows that the adsorption equilibrium data of phenolic centration of component i, respectively. The parameters q_{m,i} and compounds are well fitted by the Langmuir isotherm model and K_{L,i} are the adsorption capacity of component i in the adsorbent the correlation coefficients for epicatechin and catechin are 0.97 and adsorption affinity of component i, respectively. All these and 0.94, respectively. The maximum adsorption capacity (q_m) of parameters are obtained from single adsorption isotherms. catechin (73.56 mg/g) obtained by fitting calculation using Sigma- In a multi-component system, competition for vacant adsorp- Plot 12.0 is higher

than that of epicatechin (29.04 mg/g). The adsorption sites on the surface of adsorbent occur not only between adsorption affinity parameter (KL) of catechin and epicatechin on the same molecules but also between different adsorbate species. The adsorption constants are 0.1257 L/mg and 0.041 L/mg, respectively. Higher adsorption affinity parameter of catechin than epicatechin may act as the controlling factor for the adsorption process [16,23] and such phenomenon is not captured by ordinary Langmuir model. Recently, Kurniawan et al. [23] the carbon. and Soetaredjo et al. [16] modified the extended Langmuir model. Industrial Adsorption process usually involves multi-component system by including fractional surface loading or coverage [23] and selection, therefore, information on adsorption equilibria in multi-component system is required for engineering design of the adsorption compounds, the fractional surface loading approach by Kurniawan process. Compare to the single component adsorption, reliable et al. [23] is more

9 suitable to represent the binary adsorption data experimental

multi-component adsorption isotherm data are more of this study. Since

30 competition between catechin and epicatechin molecules for vacant adsorption sites in the activated carbon occurred during the

adsorption process, catechin and epicatechin were adsorbed in certain amount, and the adsorption capacity of the activated for both adsorbates (binary) can be written as $q_m \cdot \frac{h_1}{h_1 + h_2}$ and $q_m \cdot \frac{h_2}{h_1 + h_2}$. Here h_1 and h_2 are the fractional loading of catechin and epicatechin, respectively, and $h_1 + h_2 = 1$. The adsorption affinity parameter in Langmuir equation represents

9 how strong an adsorbate molecule is attracted onto the surface

of the adsorbent. In a binary system, the competition between different molecules weakens the adsorption potential

29 of the adsorbate toward the surface of adsorbent, therefore, the

affinity parameter of binary compounds can be written as: $KL_1 \cdot \frac{h_1}{h_1 + h_2}$ and $KL_2 \cdot \frac{h_2}{h_1 + h_2}$ where KL_1 and KL_2 are the affinity parameter of catechin and epicatechin in binary system, respectively. The final expressions of extended Langmuir model with inclusion of fractional loading concept are $q_e \cdot \frac{h_1}{h_1 + h_2}$ and $q_e \cdot \frac{h_2}{h_1 + h_2}$. The applicability of the modified extended Langmuir model was tested against the liquid phase binary adsorption data of catechin and epicatechin. The parameter δ_1 and δ_2 were obtained by the simultaneously fitting of Eqs. (9) and (10) using

34 nonlinear least square method by minimizing the following

objective function $\sum_{i=1}^n (q_{exp} - q_{cal})^2$ where q_{exp} and q_{cal} are the actual amount and the calculated values of catechin and epicatechin adsorbed by activated carbon, respectively;

34 **n is the total number of experimental data** used in **the fitting**

procedure. Fig. 2 shows the 3D plots of binary adsorption of catechin and epicatechin at various initial concentrations and the model calculated by Eqs. (9) and (10). The fitted parameters h_1 and h_2 with calculate values of $q_m(\text{bin})$, $KL_1(\text{bin})$, and $KL_2(\text{bin})$ are summarized in Table 1. In this work an improvement was made in the fitting procedure of modified extended Langmuir model by Kurniawan et al. [23]. Both Eqs. (9) and (10) were simultaneously fitted by nonlinear least square method with h_1 and h_2 as the fitted parameters, with boundary condition of the fitting procedure as $h_1 + h_2 = 1$. It is obvious that this fitting procedure

28 **improved the applicability of the modified extended Langmuir model** in correlating **the binary experimental data of**

catechin and epicatechin as indicated in Fig. 2. The maximum amount of adsorbates adsorbed in binary system ($q_m(\text{bin})$) was lower than in single system (q_m) due to the effect of molecular competition between both adsorbate molecules. This competition was accommodated in the modified extended Langmuir by taking into account of fractional loading coverage. The fractional coverage of catechin (h_1) and the maximum amount adsorbates adsorbed ($q_m(\text{bin})$) decreased with increasing initial epicatechin concentration as shown in Table 1. This result is consistent with the adsorption experimental data of catechin and epicatechin modified extended Langmuir model. (a) $q_m(\text{bin})$ (mg/g) vs C_0 (mg/L) for catechin and epicatechin. (b) h_1 vs C_0 (mg/L) for catechin and epicatechin. Fig. 2. 3D plot of binary adsorption data and modified extended Langmuir model. (a)

18 **Catechin 100 mg/L + epicatechin 100 mg/L**, (b) catechin **100 mg/L + epicatechin 150 mg/L**.

tent with adsorption characteristics in single adsorption system that the adsorption capacity of epicatechin is less than that of catechin, therefore increasing the concentration of epicatechin in binary system caused the interaction of adsorbate molecules to become even higher and reduce the value of $q_m(\text{bin})$. 3.3. Adsorption kinetics The effective design of an adsorption system requires information about the rate at which the organic compounds removal takes place in a given sorbent/solution system. Several adsorption kinetic

13 **models have been developed to describe the adsorption kinetics** in liquid phase. **The**

development of adsorption kinetic models mostly is based on fundamental approach to interfacial kinetics and its modifications [24]. Several models which have been developed 410 F.E. Soetaredjo et al. / Chemical Engineering Journal 231 (2013) 406–413 Table 1 Fitted and calculated parameters for catechin and epicatechin adsorption at several initial concentrations. Phenolic compounds mixture Fitted parameter Calculate parameter R^2 h_1 h_2 $q_m(\text{bin})$, mg/g $KL_1(\text{bin})$, L/mg $KL_2(\text{bin})$, L/mg

18 **Catechin 100 mg/ L + Epicatechin 100 mg/L** Catechin **100 mg/ L** + Epicatechin **150 mg/L**

0.632 0.368 0.524 0.476 57.1 0.0699 52.2 0.0499 0.0075 0.0139 0.998 0.997 based on this approach are Langmuir kinetics, first and second order, reversible first and second order, and irreversible reactions [24]. Another approach which is used to develop the kinetic models is based on assuming that surface reaction is the rate controlling step in sorption. Many simple and compact models have been proposed based on this approach and the famous and

39 **widely used models** are **pseudo-first** (Lagergren) **and pseudo-second order** [24]. **Pseudo-first order model** which is also known as

Lagergren equation is the earliest empirical model describing the rate of adsorption in a liquid phase system [25]. The differential form of this model is $\frac{dq}{dt} = k_1(q_e - q)$ The

29 **integration of Eq. (12) with initial** conditions: **$q(t) = 0$ at $t = 0$**

gives $q = q_e(1 - \exp(-k_1 t))$ where $q(t)$

33 **is the amount of adsorbate adsorbed at time t (mg/g), and k_1 is the rate constant (1/min). Pseudo-** second **order**

was first proposed by Blanchard et al. [26]. This model assumes that direct adsorption/desorption (seen as chemical reaction) is the rate controlling step in the overall sorption kinetics [24], and has the differential form as $\frac{dq}{dt} = k_2(q_e - q)^2$ Here

3 **k_2 is rate constant of pseudo-second order (g/mg min).** Inte- gration of **Eq.** (14) gives **the**

following result $\frac{q}{q_e(2q_e - q)} = k_2 t$ To obtain the parameters of pseudo-first and pseudo-second order equations, non-linear least square method was employed by fitting the kinetic experimental data with Eqs. (13) and (15) and the results are summarized in Table 2. Figs. 3 and 4 present the plots of catechin and epicatechin uptake as the function of time, respectively. The symbols represent the kinetic experimental data while the solid lines represent the pseudo-first and pseudo-second order kinetic models. As shown in Table 2, the values of R^2 obtained from the fitting of kinetic experimental data using both models were similar, however the

16 **pseudo-second order** kinetic **model represents kinetic** experimental **data better than the** pseu- do **-first order** since **the**

predicted amount of adsorbate adsorbed in equilibrium condition was closer to that of the measured one. This is confirmed by the fact that from the values of fitting parameter q_e of pseudo-second order model

agree with the values of q_e from kinetic experimental data. As mentioned by Plazinski et al. [24] that one of the potential advantages of using pseudo-second order kinetic model to correlate the adsorption kinetic experimental data

13 is its small sensitivity to the influence of the random experimental error,

especially for the fitting parameter q_e . Another advantage of using

13 the pseudo-second order is that the value of fitted parameter

q_e can be obtained from the independent plot of kinetic adsorption experimental data [24,27]. In many adsorption kinetic studies [24,27–29], failure of the pseudo-first order in correlating adsorption kinetic experimental data is due to the theoretical interpretations of this equation. In pseudo-first order it is based on the common assumption that the overall sorption process is controlled by the rate of adsorption/desorption processes which is seen as a chemical reaction on the surface of adsorbent [24]. However in many cases, the rate of diffusion of solute into interior of the adsorbent also controls the adsorption mechanism and this phenomenon is not captured by the pseudo-first order model. Table 2, Figs. 3 and 4 clearly indicate that the parameter k_2 plays a role as time scaling factor. At high k_2 , shorter time is required to reach equilibrium. As can be seen in Table 2, k_2 decreased with increasing initial concentration. A simple correlation between of k_2 and initial concentration of the adsorbate (C_0) is given as follows: $k_2 = \frac{1}{a} \exp(-bC_0)$ where k_2, a and b are empirical constants. Plots of C_0 versus k_2 are given in Fig. 5. The fitted empirical constants for catechin are $k_2,0 = 0.0013$ g/mg min, $a = 0.0194$ g/mg min, and $b = 0.0092$ L/mg, while for epicatechin are $k_2,0 = 0.0004$ g/mg min, $a = 0.0147$ g/mg min, and $b = 0.0207$ L/mg. Table 2 Fitted parameters of pseudo-first and pseudo-second order of catechin and epicatechin adsorption onto activated carbon NORIT nrs carbon E_a 0.5–1.5. Phenolic compound C_0 ,

19 mg/L q_e , mg/g, experiment Pseudo-first order Pseudo-second order k_1 , 1
/min q_e , mg/g R^2 k_2 , g/mg min q_e , mg/g R^2

Catechin 101.3 122.5 139.8 158.7 203.6 251.9 Epicatechin 100.5 118.4 141.2 161.5 201.7 252.3 43.3 43.9
44.5 44.7 45.3 46.1 0.0631 0.0471 0.0441 0.0337 0.0275 0.0185 21.2 21.9 23.2 25.4 28.2 28.9 0.1037
0.1096 0.0974 0.0795 0.0708 0.0656 39.1 38.6 38.3 38.2 35.6 33.6 0.996 0.997 0.998 0.998 0.996 0.994
19.6 22.2 21.1 22.5 25.4 26.5 0.993 0.996 0.996 0.995 0.994 0.995 0.0022 0.0014 0.0013 0.0008 0.0007
0.0004 43.4 44.3 44.3 46.1 44.2 45.1 0.0086 0.0083 0.0069 0.0053 0.0040 0.0034 20.9 23.6 22.8 24.5 27.9
29.3 0.997 0.998 0.999 0.998 0.997 0.996 0.994 0.998 0.996 0.996 0.997 0.995 50 (a) 40 qt (mg.g-1) 30 20
10 0 0 20 40 60 80 100 120 140 160 T (min) 50 (b) 40 qt (mg.g-1) 30 20 10 0 0 20 40 60 80 100 120 140
160 T (min) Fig. 3. Kinetic adsorption experimental data of catechin onto activated carbon NORIT nrs carbon E_a 0.5–1.5 at various initial concentration and plots of

27 (a) pseudo- first order, and (b) pseudo-second order. 3.4. Recover of

phenolic compounds in sago waste effluent To verify the potential application of recovering phenolic compounds from sago waste effluent by adsorption, adsorption experiments using real sago waste effluent were also conducted. The adsorption of catechin and epicatechin from real sago waste effluent was conducted isothermally at 303.15 K under static condition. Prior to the adsorption experiment, the sago waste effluent was filtered to remove solid particles. The initial concentration of catechin and epicatechin

in sago waste effluent was 321.8 mg/L and 158.2 mg/L, respectively. The recovery of catechin and epicatechin in sago waste effluent is

32 **given in Table 3. From Table 3 it can be seen that the** recovery of

both phenolic compounds increases with increasing activated carbon mass due to the increase of adsorption active sites. At an adsorbent dose of 2%, the

4 **recovery of catechin and epicatechin from sago waste effluent**

was 85% and 53%, respectively. It can be seen that the amount of phenolic compounds adsorbed by the activated carbon from real waste sago effluent (Table 3) is less than that of the synthetic one (Fig. 2). In waste sago effluent, in addition to catechin and epicatechin other compounds such as protein and soluble starch are also present which compete for adsorption active sites

22 **resulting in the** observed decrease **of adsorption capacity**

toward catechin and epicatechin.

22 **Reusability of adsorbent is crucial in industrial** application. **The**

recovery of catechin and epicatechin from activated carbon was performed by supercritical CO₂ at 313.15 K and 26 MPa under static and dynamic conditions. Details about the supercritical system can be found elsewhere [30]. At static condition, the recovery of catechin from activated carbon was only 1.2% while for epicatechin it was 14.5%. This low recovery is consistent with the solubility 30 (a) 25 20 qt

12 **(mg.g⁻¹) 15 10 5 0 0 20 40 60 80 100** 120 140 160 **T (min)**

30 (b) 25 20 qt

12 **(mg.g⁻¹) 15 10 5 0 0 20 40 60 80 100** 120 140 160 **T (min) Fig. 4.**

Kinetic adsorption experimental data of epicatechin onto activated carbon NORIT nrs carbon Ea 0.5–1.5 at various initial concentration and plots of

27 **(a) pseudo- first order, and (b) pseudo-second order.** data of

both phenolic compounds in supercritical CO₂ under static condition [30,31]. In order to increase the recovery of catechin and epicatechin in supercritical CO₂, dynamic extraction experiments were also conducted with flow rates of CO₂ from 0.1 to 1 kg/h. By increasing CO₂ flow rate from 0 kg/h (static mode) to 0.1 kg/h, the recovery increased from 1.2% to 2.3% and from 14.5% to 27% for catechin and epicatechin, respectively. Maximum recovery (catechin 9.7% and epicatechin 87.3%) was achieved

38at a CO₂ flow rate of 0.7 kg/h.

Further increase of CO₂ flow rate to 1 kg/h had negligible effect on the recovery both compounds. 0.010 0.008 Catechin Epicatechin k₂, g/mg.min 0.006 0.004 0.002 0.000 80 100 120 140 160 180 200 220 240 260 Initial concentration, mg/L Fig. 5. Correlation between Co and k₂. 412 F.E. Soetaredjo et al. / Chemical Engineering Journal 231 (2013) 406–413 Table 3 Adsorption of catechin and epicatechin from sago waste effluent. Co,1 catechin, mg/ Co,2 epicatechin, mg/ m, mass of activated carbon, L L g Ce,1, mg/ Ce,2, mg/ L L qe,1, mg/ qe,2, mg/ % Recovery g g catechin % Recovery epicatechin 321.8 158.2 0.112 0.483 0.689 0.921 1.104 1.328 1.573 1.921 274.4 143.3 141.2 113.8 110.3 103.8 99.8 100.2 95.5 96.4 85.4 93.1 71.2 87.4 48.3 74.2 42.3 13.3 37.4 9.2 30.7 7.9 24.1 6.3 20.5 5.6 17.8 4.9 15.9 4.5 14.2 4.3 14.7 9.4 56.1 28.1 65.7 34.4 68.9 36.7 70.3 39.1 73.5 41.1 77.7 44.7 84.8 52.3 Note: volume of sago waste effluent 100 mL. Low recovery of phenolic compound, especially catechin is caused by the difference in the polarity between the solute (phenolic compound) and the solvent (supercritical CO₂). Catechin is highly polar while supercritical CO₂ is non-polar. In order to increase the polarity of supercritical CO₂, ethanol as co-solvent (5–25%) was used in this study. By adding 5% ethanol

38at a CO₂ flow rate of 0.7 kg/h,

the recovery of catechin increased almost two fold (from 9.7% to 18.2%), and while for epicatechin the recovery reached 95.7%. Increasing the percentage of co-solvent increased the recovery of both compounds. For epicatechin the maximum recovery was 99.2% at 20% of co-solvent. Further increase in co-solvent concentration had negligible effect on the recovery of epicatechin. The un-recovered epicatechin may be strongly bound on carbon surface or epicatechin molecule may be adsorbed inside small or ultra micropore so it could not be desorbed by supercritical CO₂. At 30% co-solvent, the recovery of catechin was 93.1%. Since further increase of co-solvent concentration changed supercritical condition into subcritical, therefore maceration (303.15 K, 24 h) was used to obtain maximum removal of catechin (97.8%) and epicatechin (99.4%). 4. Conclusion The

4recovery of catechin and epicatechin from sago waste

effluent by adsorption onto activated carbon followed by supercritical CO₂ desorption was demonstrated in this work. To study the kinetic and equilibria of adsorption, synthetic aqueous catechin and epicatechin solution was used instead of sago waste effluent. The adsorption equilibria and kinetic were obtained isothermally at 303.15 K in static mode condition. The

37adsorption equilibria of single and binary system

were modeled using Langmuir and modified extended Langmuir, respectively. It was found that both of

22models could represent the experimental data well. In adsorption kinetic experiments, the

well-known pseudo-first and pseudo-second models were employed to correlate the experimental data. The pseudo-second

9model gave better representation of experimental data than the pseudo-first model. The effect **of** initial concentration on **the**

value of fitting parameters k_1 and k_2 was also studied. The results indicate that k_2 plays a role as time scaling factor, at high k_2 , shorter time is required to reach equilibrium. Adsorption using real sago waste effluent followed by supercritical CO_2 extraction indicates that this method has potential application in recovery catechin and epicatechin in industrial scale. Ethanol as co-solvent was required to obtain high recovery of catechin from the activated carbon. References [1] M.A. Soobrattee, V.S. Neerghen, A. Luximon-Ramma, O.I. Aruoma, T. Bahorun, Phenolics as potential antioxidant therapeutic agents: mechanism and actions, *Mutat. Res.* 579 (2005) 200–213. [2] L. Yu, K. Zhou, Antioxidant properties of bran extracts from “platte” wheat grown at different locations, *Food Chem.* 90 (2004) 311–316. [3] P. Vitaglione, A. Napolitano, V. Fogliano, Cereal dietary fibre: a natural functional ingredient to deliver phenolic compounds into the gut, *Trends Food Sci. Technol.* 19 (2008) 451–463. [4] A.R. Rechner, M.A. Smith, G. Kuhnle, G.R. Gibson, E.S. Debnam, S.K.S. Srari, K.P. Moore, C.A. Rice-Evans, Colonic metabolism of dietary polyphenols: influence of structure on microbial fermentation products, *Free Radical Biol. Med.* 36 (2004) 212–225. [5] A.R. Rechner, G. Kuhnle, P. Bremner, G.P. Hubbard, K.P. Moore, C.A. Rice-Evans, The metabolic fate of dietary polyphenols in humans, *Free Radical Biol. Med.* 33 (2002) 220–235. [6] O. Pourali, F.S. Asghari, H. Yoshida, Production of phenolic compounds from rice bran biomass under subcritical water conditions, *Chem. Eng. J.* 160 (2010) 259–266. [7] H. Epstein, Cosmeceuticals and polyphenols, *Clin. Dermatol.* 27 (2009) 475–478. [8] S.M. Anthonysamy, N.B. Saari, K. Muhammad, F.A. Bakar, Browning of sago (*Metroxylon sago*) pith slurry influenced by holding time, pH and temperature, *J. Food Biochem.* 28 (2004) 91–99. [9] A.T. Pei-Lang, A.M.D. Mohamed, A.A. Karim, Sago starch and composition of associated components in palms of different growth stages, *Carbohydr. Polym.* 63 (2006) 283–286. [10] M. Ahmaruzzaman, Adsorption of phenolic compounds on low-cost adsorbents: a review, *Adv. Colloid Interfaces* 143 (2008) 48–67. [11] K. Baransi, Y. Dubowski, I. Sabbah, Synergetic effect between photocatalytic degradation and adsorption processes on the removal of phenolic compounds from olive mill wastewater, *Water Res.* 46 (2012). [12] A. Dabrowski, P. Podkoscielny, Z. Hubicki, M. Barczak, Adsorption of phenolic compounds by activated carbon – a critical review, *Chemosphere* 58 (2005) 1049–1070. [13] J. Qi, Y. Guo, H. Xu, Adsorption of phenolic compounds on micro- and mesoporous rice husk-based active carbons, *Mater. Chem. Phys.* 87 (2004) 96–101. [14] C. Michailof, G.G. Stavropoulos, C. Panayiotou, Enhanced adsorption of phenolic compounds, commonly encountered in olive mill wastewaters, on olive husk derived activated carbons, *Bioresour. Technol.* 99 (2008) 6400–6408. [15] M.L. Soto, A. Moure, H. Dominguez, J.C. Parajo, Recovery, concentration and purification of phenolic compounds by adsorption: a review, *J. Food Eng.* 105 (2011) 1–27. [16] F.E. Soetaredjo, A. Kurniawan, L.K. Ong, S. Ismadji, Incorporation of selectivity factor in modeling binary component adsorption isotherms for heavy metals-biomass system, *Chem. Eng. J.* 219 (2013) 137–148. [17] C.H. Giles, D. Smith, A. Huitson, A general treatment and classification of the solute adsorption, I., *Theor. J. Colloid Interface Sci.* 47 (1974) 755–765. [18] C.H. Giles, A.P. D’Silva, I.A. Easton, A general treatment and classification of the solute adsorption isotherm, II, *J. Colloid Interface Sci.* 47 (1974) 766–778. [19] Y. Ding, D. Jing, H. Gong, L. Zhou, X. Yang, Biosorption of aquatic cadmium (II) by unmodified rice straw, *Bioresour. Technol.* 114 (2012) 20–25. [20] H. Muhamad, H. Doan, A. Lohi, Batch and continuous fixed-bed column biosorption of Cd^{2+} and Cu^{2+} , *Chem. Eng. J.* 158 (2010) 369–377. [21] C.G. Rocha, D.A.M. Zaia, R.V.d.S. Alfaya, A.A.d.S. Alfaya, Use of rice straw as biosorbent for removal of $\text{Cu}(\text{II})$, $\text{Zn}(\text{II})$, $\text{Cd}(\text{II})$ and $\text{Hg}(\text{II})$ ions in industrial effluents, *J. Hazard. Mater.* 166 (2009) 383–388. [22] K.K.H. Choy, J.F. Porter, G. McKay, Langmuir isotherm models applied to the multicomponent sorption of acid dyes from effluent onto activated carbon, *J. Chem. Eng. Data* 45 (2000) 575–584. [23] A. Kurniawan, H. Sutiono, N. Indraswati, S. Ismadji, Removal basic dyes in binary system by adsorption using rarasaponin–bentonite: revisited extended Langmuir model, *Chem. Eng. J.* 189–190 (2012) 264–274. [24] W. Plazinski, W. Rudzinski, A. Plazinska, Theoretical models of sorption kinetics including a surface reaction mechanism: a review, *Adv. Colloid Interfaces* 152 (2009) 2–13. [25] S. Lagergren, Zur

theorie der sogenannten adsorption geloster stoffe Kungliga Svenska Vetenskapsakademiens, Handlingar 24 (1898) 1–39. [26] G. Blanchard, M. Maunaye, G. Martin, Removal of heavy metals from waters by means of natural zeolites, *Water Res.* 18 (1984) 1501–1507. [27] A.N. Kosasih, J. Febrianto, J. Sunarso, Y.-H. Ju, N. Indraswati, S. Ismadji, Sequestering of Cu(II) from aqueous solution using cassava peel (*Manihot esculenta*), *J. Hazard. Mater.* 180 (2010) 366–374. [28] A. Kurniawan, H. Sutiono, Y.-H. Ju, F.E. Soetaredjo, A. Ayucitra, A. Yudha, S. Ismadji, Utilisation of rarasaponin natural surfactant for organo-bentonite preparation: application for methylene blue removal from aqueous effluent, *Micropor. Mesopor. Mater.* 142 (2011) 184–193. [29] A. Kurniawan, S. Ismadji, Potential utilization of *Jatropha curcas* L press cake residue as new precursor for activated carbon preparation: application in methylene blue removal from aqueous solution, *J. Taiwan Inst. Chem. Eng.* 42 (2011) 826–836. [30] F.E. Soetaredjo, S. Ismadji, Y.-H. Ju, Measurement and modeling of epicatechin solubility in supercritical carbon dioxide fluid, *Fluid Phase Equilib.* 340 (2013) 7–10. [31] F.E. Soetaredjo, S. Ismadji, M.J. Liauw, Y.-H. Ju, Catechin sublimation pressure and solubility in supercritical carbon dioxide, *Fluid Phase Equilib.* (2013). submitted for publication.

10Author's personal copy Author's personal copy F. E. Soetaredjo et al.

/ *Chemical Engineering Journal* 231 (2013) 406–413 407

10Author's personal copy Author's personal copy F. E. Soetaredjo et al.

/ *Chemical Engineering Journal* 231 (2013) 406–413 409

10Author's personal copy Author's personal copy F. E. Soetaredjo et al.

/ *Chemical Engineering Journal* 231 (2013) 406–413 411 Author's personal copy Author's personal copy F.E. Soetaredjo et al. / *Chemical Engineering Journal* 231 (2013) 406–413 413